Thermal niche predicts recent changes in range size for bird species

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Abstract

Species distributions are strongly affected by climate, and climate change is largely impacting on species and populations. Thermal niches have been widely used as proxies for estimating thermal sensitivity of species and have been frequently related to community composition, population trends and latitudinal/elevational shifts in distribution. To our knowledge, no work has yet explored the relationship between thermal niche and change in range size (changes in the number of occupied spatial units over time) in bird species. In this paper, we related a 30 years change in range size to Species Thermal Index (STI: average temperature at occurrence sites) and to other factors (i.e. birds’ associated habitats, body mass, hunting status) potentially affecting bird populations/range size. We analysed trends of breeding birds range in Italy for a suite of poorly studied cold-adapted animals potentially sensitive to global warming, and a related group of control species taxonomically similar and with comparable mass but mainly occurring at lower/warmer sites. We found a strong positive correlation between change in range size and STI, confirming that recent climatic warming has favoured species of warmer climates and adversely affected species occupying colder areas. A model including STI and birds’ associated habitats was only marginally less supported, with forest species performing better than alpine open habitat and agricultural ones. In line with previous works highlighting effects of recent climate change on community composition, species' population trends and on poleward/upward distributional shifts, we found STI to be the most important predictor of change in range size variation in breeding birds.

**Keywords:** change in range size; climate change; cold-dwelling; warm-dwelling; Alps.
Introduction

Climate is a major factor determining a species’ distribution as this largely depends on species-specific physiological tolerance to temperature (Woodward 1987; Hoffman & Parson, 1997; Kreyling et al. 2015), precipitation (Tingley et al. 2009; Toledo et al. 2012; Illán et al. 2014) and on climate-driven habitat characteristics (Deutsch et al. 2015). Recent changes in climate have induced range shifts of animal species towards higher latitudes and elevations (Thomas & Lennon 1999; Parmesan & Yohe 2003; Zuckerberg et al. 2009; Gillings et al. 2015), altered species’ phenology (Roy & Sparks 2000; Fitter & Fitter 2002) and affected population dynamics (Thompson & Ollason 2001; Stephens et al. 2016). Direct and indirect effects of climate change are expected to aggravate the conservation status of many organisms in the current century, with numerous species predicted to become endangered (Parmesan et al. 1999; Thomas et al. 2004; Cahill et al. 2013).

Many studies investigating climate change effects on animal populations have focused on bird species as they are considered good indicators due to their ability to respond rapidly to environmental changes and spatially track suitable climatic conditions (Chen et al. 2011; Maggini et al. 2011). Studies investigating wide-scale effects of climate change on bird populations have commonly used the Species Thermal Index (STI), i.e. the average temperature across a given species’ distribution, as a proxy for estimating the thermal sensitivity of a species (e.g. Gaüzère et al. 2015). These kinds of studies typically modelled population trends (changes of population units over time) as a function of climate change (e.g. Jiguet et al. 2010). Other studies evaluated changes in the Community Thermal Index (CTI, i.e. the average temperature for a community of species) over time (e.g. Gaüzère et al. 2015; Tayleur et al. 2016). Generally, those studies reported that population trend varied according to species’ STI and that CTI increased coherently with increased temperature (Tingley et al. 2009), displaying consistent latitudinal variation (Devictor et al. 2008). However, to the best
of our knowledge, until now no study has yet considered the net change in species’ range size in a given geographical area over time as a function of the mean temperature experienced by the species across their distributional range. Changes in range size (changes in the number of occupied spatial units over time in a geographical region) related to the climatic sensitivity of the species would signal a very important effect of climate, different than the reported change in population growth (e.g. Jiguet et al. 2010) and community structure (e.g. Gaüzère et al. 2015) in a given area, and even different from a range shift which does not affect range size. This is particularly relevant for cold-dwelling species like mountain birds, which are predicted to be affected by global warming more than warm-dwelling species (Araújo et al. 2011; Pearce-Higgins et al. 2015; Tayleur et al. 2016). Compared to birds in other habitats, mountain birds have been relatively poorly studied, largely due to the logistical problems in surveying in such challenging climatic and topographic environments (Chamberlain et al. 2012; Scridel 2014), but are particularly at risk because of global warming as species are often confined to discrete habitat isolates and can not perform a latitudinal track of suitable climate (Fjeldså et al. 2002; Bech et al. 2009).

Here we investigate the relationship between change in range size and STI for a set of cold dwelling birds and for their closest (non cold-dwelling) relatives breeding in Italy, whilst accounting for other important factors known to potentially affect bird trends such as species broad habitat association (e.g. Chamberlain et al. 2013), hunting status (e.g. Sandercock et al. 2011) and body mass (e.g. Brommer 2008). Moreover, various studies have shown that population trends (Thomas 2008) and extinction risks (Purvis 2008) are not randomly distributed in respect of phylogeny: species that share long evolutionary history are more likely to exhibit similar responses then evolutionary distant ones. Therefore to provide an unbiased assessment of climate change effects we used a phylogenetic comparative approach, which accounts for evolutionary relatedness between species.
Methods

Model system

We worked with the reported change in range size in the last 30 years for breeding bird species in Italy. Italy represents a good study model to investigate the relationship between birds' climatic niche and the variation of their range size trends because i) it encompasses a wide latitudinal and elevational gradient, ii) it comprises different biogeographical regions with different species assemblages (i.e. from Mediterranean islands to high mountains), iii) it includes one third of the Alps, a mountain chain for which strong impacts of climate change have been already observed (Sergio 2003; Maggini et al. 2011; Pernollet et al. 2015) and may act synergically with changes in land use (e.g. Laiolo et al. 2004). Europe has recently experienced a linear increase in average temperatures from the 1980s onwards, by over 0.9°C (KNMI 2007), but in the Alps the rate of warming has been double the global average (Brunetti et al. 2009) and rising temperatures, higher snow lines and lower snowfalls are predicted to continue (EEA 2010). Adverse effects are therefore expected for these regions hosting many geographically isolated species, often being glacial relicts occurring at the edge of their ecological and climatic niche (e.g. boreo-alpine species and other species found only in high mountains) and often predicted to undergo range contractions and/or population decline as a response to climate change (Maggini et al. 2011; Chamberlain et al. 2013; Viterbi et al. 2013; Braunisch et al. 2014; Brambilla et al. 2015; Pernollet et al. 2015; Brambilla et al. 2016; Brambilla et al. 2017a). To understand whether changes in range size differ according to different STI, we investigated changes in range for breeding bird species by considering a set of species comprising closely related taxa with largely different STI, in order to get an indirect measure of the species' sensitivity to climate change. Firstly, we chose bird species displaying in Italy a breeding distribution strictly related to the main mountains (i.e. Alps and Apennines; Nardelli et al. 2015). We included boreo-alpine species (rock ptarmigan *Lagopus*...
muta, pygmy owl *Glaucidium passerinum*, boreal owl *Aegolius funereus*), species occurring in southern European mountains (alpine accentor *Prunella collaris*, yellow-billed chough *Pyrrhocorax graculus*, white-winged snowfinch *Montifringilla nivalis*) and other species that in southern Europe and in the Alps occur only or predominantly on mountains or in relatively cold areas: hazel grouse *Bonasa bonasia*, black grouse *Lyrurus tetrix*, capercaillie *Tetrao urogallus*, nutcracker *Nucifraga caryocatactes*, water pipit *Anthus spinoletta*, citril finch *Carduelis citrinella*, common redpoll *Carduelis flammea*, ring ouzel *Turdus torquatus*, wallcreeper *Tichodroma muraria*, three-toed woodpecker *Picoides tridactylus*, black woodpecker *Dryocopus martius*, grey-headed woodpecker *Picus canus*, willow tit *Parus montanus* (although the latter three species occur broadly in lowland habitats in central-northern Europe, in the Alps and in southern Europe they are mostly associated with relatively cold areas). Secondly, for each of the above mentioned taxa (*n*=19), whenever possible a control species was chosen (*n*=19), taking the one most closely related according to Roquet et al. (2014) and generally associated with lower elevation in order to have a balanced sample (Table 1). With such *ad hoc* species selection, we ensured a phylogenetically balanced sample of species, the inclusion of cold-dwelling taxa and of other not specialized for cold environments, the focus on species with similar movement strategies (all species apart tree pipit *Anthus trivialis* being resident or short-distance migrants), the exclusion of species for which change in range size may be due mostly to improved knowledge or reintroduction/restocking projects carried out between the two reference periods (as actually occurred for some species; Nardelli et al. 2015).
To quantify STI for each species we gathered breeding distribution data available from the EIONET portal (http://www.eionet.europa.eu), which provided presence/absence data at a 10km x 10km resolution for 23 European countries (i.e. UK, France, Italy, Slovenia, Spain, Portugal, Austria, Denmark, Sweden, Finland, Greece, Belgium, Gibraltar, Slovakia, Malta, Lithuania, Latvia, Ireland, Hungary, Germany, Estonia, Romania, Czech). For each species, we merged all available data to create a large-scale breeding distribution map using a geographic information system (GIS) software (QGIS; Quantum GIS Development Team 2016; GRASS; GRASS Development Team 2015). We then coupled the above-mentioned species’ digitized breeding distribution with 30-arc seconds resolution data describing the mean annual temperature (downloaded from Worldclim database version 1.4; http://www.worldclim.org; Hijmans et al. 2005). STI were consequently calculated by averaging mean, median, minimum and maximum temperature (in °C) experienced by each species during the breeding season across its distribution range. To date, many studies have used a variety of thermal indexes to evaluate potential responses of bird communities to climate change ranging from thermal maximum and minimum (mean of hottest/coldest 5% cells; e.g. Jiguet et al. 2010), thermal range (difference between thermal maximum & minimum; e.g. Jiguet et al. 2010), seasonal thermal average (mean temperatures during breeding period; e.g. Devictor et al. 2008), coldest-month mean temperature (e.g. Green et al. 2008), annual temperature sum above 5°C (e.g. Green et al. 2008). To minimize collinearity, we used variance inflation factors (VIFs) on our four temperature predictors and highly collinear variables (VIF > 5) were omitted following Zuur et al. (2009). There were high levels of collinearity between these variables and only mean annual temperature was used considered the be the most representative estimate for the thermal niche of the study species,
due to the largely non-migratory status of our target species (most species are resident, and only tree pipit *Anthus trivialis* is a long-distance migrant).

To estimate change in range size, we used long-term bird distribution trends from Nardelli et al. (2015), which compared occupied vs unoccupied cells of breeding species in Italy between the 1980s (1983-1986, according to data available for each species) and 2012, therefore calculating a net change of range size (Table 1). Species ranges were estimated by collating all available data, such as regional and local atlases, published reports, papers, monitoring schemes (Nardelli et al. 2015). Given that most of the sources used to define species range were general and not species-specific, the potential effect of confounding factors such as an increase of knowledge should affect in the same way all the species, thus it is unlikely to produce any bias in our analyses. The only partial exception to this pattern is represented by an increased effort at high elevation in the common birds monitoring scheme (MITO 2000 project Pan-European Common Bird Monitoring Scheme). We stress that species’ net changes in range size reported in Nardelli et al. (2015) are unlikely biased in respect of elevation as these are in line with various national (Fornasari et al. 2004; Gustin et al. 2010; Rete Rurale Nazionale & LIPU 2014) and international studies on population trends (i.e. Zbinden et al. 2005), which generally highlight a wide scale decline for mountain birds, particularly marked for open-habitat species.

**Statistical analysis**

To tests the hypothesis that STI is associated with changes in range size (and thus that cold-adapted species have contracted more their range size than warm-dwelling species), we fitted phylogenetic general least square models (PGLS) implemented in the package “Caper” (Orme
et al. 2013; ver.0.5.2) in the statistical environment “R” version 3.2.3 (R Core Team 2015). The response variable was the long-term trend (as percentage variation reported in Nardelli et al. 2015) in the national breeding range of the selected bird species. Explanatory variables were the STI of each species and three other factors potentially driving variation: the broad habitats to which the species is mostly associated in Italy (forest [n=16 species], alpine open [n=12] and agricultural habitats [n=10]; Cramp et al. 1977 – 1994), a categorical variable defining the prevalent hunting status in Italy (i.e. hunted/non-hunted; Art. 2, Law 157/92) and the average body mass of each species as a proxy for demographic traits (Cramp et al. 1977 – 1994; Julliard et al. 2003; Brommer et al. 2008). An interaction term between species’ associated habitat and STI was also included in order to test the hypothesis that the relationship between STI and changes in range size differs across habitats. PGLS models were chosen in order to incorporate the covariance between related birds and therefore accounting for the non-independence of data points due to common ancestry (Paradis 2014). Firstly, we built a phylogenetic tree for our selected species based on the supertree from Jetz et al. (2012) and trimmed it using the related website www.birdtree.org. A co-variance matrix was then produced calculating the branch lengths of the phylogenetic tree and fitted in the PGLS model to estimate maximum likelihood of the parameter λ and for phylogenetic signal in the model residuals (Pagel’s λ). A value of λ = 0 represents no phylogenetic signal, whilst λ = 1 means a high phylogenetic signal and therefore consistent with a Brownian motion model of trait evolution (Pagel 1999). Model selection for fixed terms was then performed by comparing Akaike’s Information Criterion corrected for small sample size (AICc; Burnham & Anderson 2002) using the dredge function in the R package ‘MuMIn’ (Barton 2015), which allows to generate mathematical models (PGLS in our case) using all combinations of variables.

Results
Our sample of 38 species covered a wide range of thermal positions (mean STI 7.6 ± 0.4 °C, range 3.1°C to 13°C, uncertainties expressed in SE units), with the ‘coldest’ species being rock ptarmigan *Lagopus muta* and the ‘warmest’ being rock sparrow *Petronia petronia*. STI values were cooler for species of alpine open habitat (mean 5.6 ± 0.4 °C) than for species of forest (mean 7.1 ± 0.6 °C) and agricultural habitat (mean 10.7 ± 0.4 °C). Changes in range size were negative for alpine open habitat species (mean % trend: -4.4 ± 2.3), positive for agricultural (12.9 ± 6.4) and, especially, for forest species (14.2 ± 5.1). Based on AICc, the most supported model explaining change in range size in bird only included STI as a fixed variable (Table 2) and explained a good proportion of the variance (adjusted-$R^2 = 0.36$). All other models with similar support ($\Delta$AICc <2) included the most supported as a nested model, thus all the other parameters may be considered uninformative. For the most supported model, Pagel’s $\lambda$ was equal to 0.48.

We thus found a positive correlation between long-term changes in range size of Italian breeding birds and the respective thermal niche (STI) of a species (Fig. 1): species which have experienced losses in their distribution were mostly associated with cold STI values. A poorly supported effect (uninformative parameters) was found for habitat and for the habitat x STI interaction, generally suggesting a more favourable change in range size for forest species than for those dwelling open habitats and, especially, agricultural ones.

**Discussion**

This work highlights for the first time the importance of thermal niche (described by means of the average temperature experienced by a species across its breeding range) as a predictor for the change in breeding range size for a suite of European birds with different thermal niches.
and inhabiting different habitats. Our work analysed changes in range size in Italy, which represents an optimal context for such an assessment, given its location at the southern boundary of Europe and the presence of wide latitudinal and elevational gradients.

Thermal niche has been frequently reported as the most important component of the climatic niche of birds and it is known to be a reliable tool to evaluate climate change effects on species dynamics (Jiguet et al. 2007; Barnagaud et al. 2012; Howard et al. 2015; Stephens et al. 2016). Our findings are consistent with previous studies reporting the importance of thermal niche in predicting bird population trends (Jiguet et al. 2010; Thaxter et al. 2010; Pearce-Higgins et al. 2015) and complement them by highlighting the relationship between thermal niche and range dynamics, in addition to the already reported link between thermal niche and population trends.

The overwhelming importance of STI in predicting changes in range size variation for breeding bird species in the last 30 years provides additional support for the strong effect of climate change on the change in the size of their breeding range, even over a relatively short timeframe. Thousands of cases of species changing their distribution in relation to climate change have been already reported (Bellard et al. 2012), mostly in the form of latitudinal and elevational shifts, which seem particularly frequent in species displaying good dispersal abilities (Parmesan 2006). Range contraction have been indeed reported for species of polar and high-mountain regions (Forero-Medina 2010). Consistently, our results further confirmed how cold-adapted species inhabiting mountain regions are at particular high risk, as their range has already contracted most likely because of climate change effects. Indeed it is likely easier for lowland species to track suitable climates by performing latitudinal shifts, whereas for mountain species, confined to discrete habitat isolates, tracking suitable climates almost invariably means contracting the relative range (Fjeldså et al. 2002; Bech et al. 2009). In fact,
also species showing no or positive changes in range size might have shifted their range, but only in the case of cold-adapted species range shifts resulted in definite range contraction. The positive correlation between changes in range size bird distribution trends and STI supported our initial expectation, suggesting that recent climatic warming has favoured species adapted to warm areas while adversely impacting on species occupying cool sites. Tayleur et al. (2016) found that changes in Community Thermal Indexes were driven by warm-dwelling species colonising new sites, whereas cold-dwelling species contracted. Pearce-Higgins et al. (2015) found that cold-adapted species experienced more negative effects of higher temperatures than species associated with warmer temperatures. In addition to thermal niche, a minor effect of habitat (an 'uninformative parameter' according to Arnold 2010) partly emerged from models, basically highlighting a positive change in size for forest species (as shown by position of e.g. woodpecker species in Fig. 1), coherent with the better conservation status shown by woodland birds at the national level, determined by an increase in both woodland cover and quality (see Brambilla et al. 2013 and references therein). The only cold-dwelling species showing a positive change in range size is three-toed woodpecker, and in general woodpeckers, tits and other typical forest-dwelling species showed positive variation of range size. Some forest species experienced more positive variation in range size than some farmland species with a higher STI (Fig. 1). However, consistent with the prominent effect of STI, even the forest species with the lowest STI values, such as boreal and pygmy owl, have experienced negative changes in their breeding distribution: these are species showing a cold STI (4.9 C° and 5.1 C° respectively) and they might be particularly sensitive to climate warming. This interpretation is in line with other studies, which predicted boreal and pygmy owl distribution to further contract according to future global warming scenarios (Brambilla et al. 2015, Brambilla et al. 2017a). In contrast, the respective control species in the same habitat (i.e. tawny owl Strix aluco) has experienced
a positive trend in the last 30 years whilst having a higher STI (9.6 °C). Tawny owl
distribution is known to be limited by climate (Vrezec & Tome 2004) and recent increases in
distribution and breeding success in Finnish populations have been attributed to climate
warming (Mikkola 1983; Solonen 2005).
Except for tree pipit, all other alpine open habitat species were associated with long-term
losses in range and low STI. For some of those species (e.g. ptarmigan) strong contractions
have also been observed in other Alpine regions (Revermann et al. 2012; Pernollet et al.
2015), with global warming effects likely to influence not just the habitat but also the
breeding mechanisms of species (Martin & Weibe 2004). Furthermore, our results agree also
with other studies, which have forecasted future losses in distribution of open habitat species
(i.e. water pipit) due to climate and to the disappearance of alpine grassland as it becomes
slowly colonised by shrub and trees (Chamberlain et al. 2013). Even if we cannot rule out the
potential impact of land use change, and in particular of land abandonment (see e.g. Brambilla
et al. 2010, 2017b), modelling results suggested an overwhelming importance of STI and thus
a major impact of climate change.
In conclusion, our work provided evidence for a direct link between species' thermal niche
and the change in range size in the last 30 years. Coherently with previous works highlighting
an effect of recent climate change on community composition, species' population trend, and
on poleward or upward distributional shift, we found climate warming to be the most likely
factor in explaining the change in range for breeding birds in Italy.

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